An exposure and risk assessment for fluoride and trace metals in black tea

Sait C. Sofuoglu* 1, Pınar Kavcar1

Izmir Institute of Technology, Department of Chemical Engineering and Environmental Research Center, Gülbahçe, Urla 35430 Izmir, Turkey

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Abstract

Exposure and associated health risks for fluoride and trace metals in black tea were estimated. Fifty participants were randomly recruited to supply samples from the tea that they drink, and self-administer a questionnaire that inquired about personal characteristics and daily tea intake. Analyzed trace metals included aluminum, arsenic, barium, cadmium, cobalt, chromium, copper, manganese, nickel, strontium, and zinc. Fluoride and four metals (Al, Cr, Mn, Ni) were detected in all samples while barium was detected only in one sample. The remaining metals were detected in >60% of the samples. Fluoride and aluminum levels in instant tea bag samples were greater than in loose tea samples (p<0.05) while the differences in elemental concentrations of loose and pot bag tea samples were not significant. Median and 90th percentile daily tea intake rates were estimated as 0.35 and 1.1 l/day, respectively. Neither fluoride nor aluminum levels in black tea were found to associate with considerable risks of fluorosis and Alzheimer’s disease, respectively. However, carcinogenic risk levels for arsenic were high; R > 1.0 × 10⁻⁶ even at the median level. According to sensitivity analysis, daily tea intake was the most influencing variable to the risk except for arsenic for which the concentration distribution was of more importance.

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1. Introduction

Elemental content of tea may have both beneficial and adverse effects on human health. While beneficial effects were investigated against cancer [1,2], vascular diseases [3], hypertension [4], and dental caries [5], tea was associated with dental [6] and skeletal [7] fluorosis, and Alzheimer’s disease [8] due to accumulation of fluoride (F) and aluminum (Al), respectively, in the plant. The impact of tea drinks on human health has been observed, and trace elemental content of tea has been investigated in Southern Asian countries where tea is a traditional drink consumed in large quantities, such as in China [9], India [10], Taiwan [11], and Tibet [6] for fluoride, and in China [12], and Thailand [13] for aluminum and other trace elements. Recently, infusion fluoride concentrations of up to 6.5 mg/l [7] and brand name bottled tea concentrations of up to 4.1 mg/l [14] were measured in the US.

Aluminum is liberated from the aluminosilicate fraction of soil clays in acidic conditions, under which F–Al complexes can be formed [12], which leads to their elevated uptake by the tea plant. Transported to the leaves, they both can be accumulated at large quantities. Shu et al. [15] reported that Al and F contents were higher in older tea leaves, which was also associated with lower tea quality [16]. Higher F levels were measured in infusions of instant black tea bags than of granular and stick shaped black tea [18]. In fact recently, skeletal fluorosis diagnosed on an American patient was found to be associated with her long-term intake of instant tea in large quantities [7]. While almost all the fluoride could be infused into water [17], transfer ratio from tea product to liquor was estimated as 0.34–0.58 for aluminum [12]. Fluoride concentration in tea infusions may be up to 6.1 mg/l for instant black tea bags [7], and 7.3 mg/l for brick tea after 6-h infusion [19]. Infusion aluminum concentrations were reported as 0.7–3.5 mg/l [12] but might be up to 6 mg/l [19].

* Corresponding author. Tel.: +90 232 750 6648; fax: +90 232 750 6645.
E-mail addresses: cemilosofuoglu@iyte.edu.tr, saitcemil@iit.edu (S.C. Sofuoglu), pinarkavcar@iyte.edu.tr (P. Kavcar).

1 Tel.: +90 232 750 6670; fax: +90 232 750 6645.
Turkish people are, also traditionally, a black tea drinking population throughout the day. Fluoride content of black tea marketed in Turkey was the subject of several investigations. While F content of tea leaves was found to range between 67 and 289 mg/kg [20,21], infusion concentrations ranged from 0.64 to 3.92 mg/l [17,20]. Content of black tea other than F was studied [22] only for cadmium and vanadium (mean concentrations of two types of tea were reported as 2.79 and 4.39 mg/kg for cadmium, and 0.65 and 2.30 mg/kg for vanadium). However, manganese, iron, copper, zinc, and nickel were measured in green tea as 1610, 342, 32, 28, and 26 mg/kg, respectively [23]. Kalayci and Somer [17] estimated a daily F intake range of 0.46–0.98 mg/day for different brands of black tea and various infusion times. Tokalioglu et al. [20] concluded that, excluding intake from other sources, high-end daily tea consumption may result in dental fluorosis.

Because Turkish people habitually drink black tea, an exposure and risk assessment for trace elements in tea was deemed necessary as this information were not available in the literature. The objectives of this study were to determine (1) fluoride and trace metal infusion concentrations, especially aluminum, in black tea consumed by Izmirians, (2) daily tea consumption rate of İzmir population, and (3) associated exposure and health risk levels.

2. Materials and methods

2.1. Questionnaire

A questionnaire was composed to determine daily tea intake. The questionnaire acquired the consumption rate in different cup types as number of glasses drunk in a day. Turkish people traditionally drink tea in small (75 ml), special tea glasses. However, two types of tea cups (100 and 150 ml) are also popular. The largest cups are in the size of 250 ml. The consumption rates were then converted to liters per day. The participants self-administered the questionnaire, and provided duplicate samples from the tea they consumed. The questionnaire also acquired information about the provided tea sample including brand of the tea and how it was consumed (by brewing in a pot, instant in a cup). Other key data obtained by the questionnaire are personal information on the participant, i.e., body weight, sex, year of birth, etc. Fifty participants were randomly recruited, taking population age distribution into account.

2.2. Sample preparation

All beakers and HDPE bottles were kept in 20% nitric acid (Merck) bath for 2 h and dried in a hood. Two grams of tea were weighed from each sample. Two hundred milliliters of boiling ultra-pure water (Millipore Elix5) were added to each sample in a beaker, and infused for 15 min. At the end of the infusion period, the tea drink was filtered into two 60-ml HDPE bottles (one for F analysis and one for trace metal analysis) and cooled to room temperature. Trace metal samples were acidified to pH < 2 with the addition of ultra pure nitric acid (Fluka). All samples were stored at +4 °C in a refrigerator until analysis; for a maximum of 2 days for F analysis.

2.3. Fluoride analysis

Fluoride analysis was performed using a Corning model 450 digital pH/ion meter in conjunction with Cole-Parmer fluoride electrode. Total ionic strength adjustment buffer (TISAB) solution was prepared. To prepare the TISAB solution 28.5 ml glacial acetic acid, 29 g NaCl, and 2 g of a chelating agent (CDTA, 1,2-cyclohexylene dinitrilo tetraacetic acid) were added to approximately 250 ml distilled water in a 500-ml beaker, and stirred to dissolve the materials. The solution was completed to 500 ml by adding 5 M sodium hydroxide, resulting in a solution with a pH of 5.0–5.5. The TISAB solution regulates the ionic strength of samples and standard solutions, and adjusts the pH. It also avoids interferences by polyvalent cations such as Al(III), Fe(III), and Si(IV) that are able to complex or precipitate with fluoride, and reduce the free fluoride in the solution [20].

A series of fluoride standards were prepared by using 10 ppm F standard solution in the range of 0–10 mg/l by diluting appropriate volumes to 50 ml. Then, the electrode was calibrated to concentrations of 0.0, 0.5, 1.0, 2.0, 5.0 and 10.0 mg/l. For every 15 ml sample, 15 ml TISAB solution was added in a 50-ml beaker. The content was stirred at medium speed for 5 min using a magnetic stirrer. The fluoride electrode was immersed in the stirred solution. When taking measurements, the electrode was remained in solution until the apparatus gave a signal and the potential (mV) was read. Calibration curve for potentiometric determination of fluoride was drawn after reading the potentials. The mean blank level was 0.163 mg/l (n = 3) for the experiment. The blank value was subtracted from sample readings.

2.4. Trace metal analysis

Inductively coupled plasma-mass spectrometry measurements were performed by a quadruple Agilent 7500ce spectrometer equipped with Octapole Reaction System (ORS). A glass concentric nebulizer (Glass Expansion, MicroMist Nebulizer) with a Peltier-controlled quartz spray chamber was used as the sample introduction system. The ShieldTorch System (STS) was used to obtain cool plasma. Calibration was performed with external standards. Instrumental operating conditions were as follows: RF generator frequency 27 MHz, power output 1500 W, argon flow rate: plasma 15 l/min, auxiliary 1 l/min, carrier 1 l/min, nebulizer 0.08 rps. Solution uptake rate 0.3 rps, interface: nickel sampler (1 mm i.d.) and skimmer (0.4 mm i.d.) cones. Data acquisition: peak hopping, dwell time 100 ms, number of replicates 3. Analytical masses: 27Al, 52Cr, 55Mn, 59Ni, 60Co, 63Cu, 66Zn, 75As, 88Sr, 111Cd, 137Ba. Ultra-pure water (Millipore Elix5), spiked with known concentrations of the 11 elements, was put through the extraction and analysis procedure for recovery assessment (n = 3). The mean recovery values were ranged from 96% for Al to 110% for Zn. Overall, recovery was 103 ± 1.7% (mean ± S.D.). Three blanks were also analyzed. The mean blank levels were <0.5% of the mean
sample concentrations for all elements, except for Cd which was 3.5%.

2.5. Exposure and risk estimation

In this study, exposure from ingestion of tea was assessed. In order to estimate the daily exposure of an individual, the US Environmental Protection Agency, USEPA [24] suggests the lifetime average daily dose (LADD) as the exposure metric. The following equation is a similar representation of daily exposure for ingestion route modified from the USEPA [25]:

\[
\text{CDI} = \frac{C \times DI}{BW},
\]

where CDI is the chronic daily intake (mg/(kg d)), C is the contaminant concentration in tea infusion (mg/l), DI is the average daily intake rate of tea (l/day), and BW is body weight (kg). Multiplication of C and DI is daily fluoride intake (mg/day). Values of these three input variables, specific to each participant, were used to estimate the subject’s individual chronic daily exposure level.

Lifetime cancer risk associated with ingestion exposure was calculated using the following equation [26,24]:

\[
R = \text{CDI} \times SF,
\]

where R is the probability of excess lifetime cancer (or simply risk), CDI is the chronic daily intake (mg/(kg day)), and SF is the slope factor of the chemical (mg/(kg day))\(^{-1}\).

The hazard quotient (HQ) was calculated to estimate non-carcinogenic risk using the following equation [27]:

\[
\text{HQ} = \frac{\text{CDI}}{\text{RfD}},
\]

where RfD is the reference dose (mg/(kg day)). An HQ value of \(\gg 1\) implies a significant risk level.

SF and RfD values employed in this study were obtained from the USEPA [28]. In addition to the individual assessment, population exposure-risk assessment was carried out for each element using Monte Carlo simulation described in the next section.

2.6. Statistical methods

Concentration data were censored for non-detects to avoid overestimation of exposure and risk. A robust method was used to censor the data. Probability distributions were fitted to the detected concentrations of contaminants with >50% detection ratio, then values were generated for the non-detects by extrapolating below the detection limit. Generated concentrations were then used in exposure and risk calculations along with the measured concentrations. Statistical analyses were performed using SPSS (Release 12.0); Monte Carlo simulations were performed using Crystal Ball (v 4.0e) software. Monte Carlo simulation is a computer-based method of analysis that uses statistical sampling techniques in obtaining a probabilistic approximation to the solution of a mathematical equation or a model [29]. Exposure and risk distributions of İzmir population were estimated using the simulated values (\(n = 10,000\)). One-way ANOVA test and t-test were used to compare the means of different groups when underlying distribution was normal. Otherwise, Kruskal–Wallis and Mann–Whitney tests were used to determine whether the infusion contaminant concentrations and risks associated with exposure to these contaminants differed across groups such as tea type and tea brand. However, the samples sizes of the subgroups were small in some instances; so the data were pooled to alleviate the effect of this drawback when it is possible. In this study, \(p\)-values < 0.05 were considered to point a significant difference between the compared groups.

3. Results and discussion

There are various sources of trace elements such as food, drinking water, etc. Some of the sources are critical for exposure to individual elements, as in the case of toothpaste, fluoridated drinking water, and tea for fluoride, and anti-acid tablets, dietary intake, tea, and alum-treated drinking water for aluminum. In this study, black tea samples were collected and infused in ultra pure water to single out the exposure from tea. Therefore, the estimated risk levels are solely for the exposure from black tea, not accounting exposures from any other sources.

3.1. Questionnaire data

Randomly recruited 50 people, ranging from 8 to 79 years old, participated in the study. The mean age of the participants was 35. Characteristics of the participants are presented in Table 1. The majority of the participants brewed loose tea in a pot with tap water, and indulged in three brands. Participant daily tea intake and body weight distributions are shown in Fig. 1. Body weights of the participants were normally distributed, which ranged from 20 to 114 kg. Daily tea intake ranged...
Fluoride was detected in all samples. Overall mean fluoride concentration was 0.68 mg/l. The concentrations were below the lowest American bottled water limit of 1.4 mg/l determined for hot climates [31]; however, 2% of the samples were greater than the Turkish limit (1 mg/l) set for non-alcoholic beverages [32]. Loose and pot bag samples had similar mean/median concentrations (0.672/0.666 mg/l vs. 0.676/0.653 mg/l, respectively). However, instant bags had a higher mean/median concentration (0.765/0.767 mg/l). The difference was not significant with t-test between loose and instant bag (p = 0.31), and between pot bag and instant bag (p = 0.35) samples. Because of the large sample size difference between loose (n = 38) and instant bag (n = 6) groups, a nonparametric test was applied. The difference between the median values of loose and instant bag samples was significant with Mann–Whitney test at p = 0.10. The difference among median concentrations of the most used three brands was not significant (p = 0.54, Kruskal–Wallis test).

Several studies were conducted on the F content of tea products sold in Turkey. Kalayci and Somer [17] measured infusion concentrations of 2.60 and 3.92 mg/l with 5 and 20 min extraction times, respectively. Water-soluble F content of Turkish teas was found to range from 55 to 127 μg/g [20]. These levels correspond to 0.55–1.27 mg/l when converted into infusion concentrations (2 g tea infused in 200 ml of water). These concentrations are similar to the levels measured in this study. Hudaykuliev et al. [21] reported a range of F content for Turkish teas as 88–289 mg/kg which corresponds to 0.8–2.6 mg/l infusion concentrations with the assumption of water-soluble fraction is 90%. Higher levels were reported from other parts of the world as 0.45–6.5 mg/l [7,10,11,18]. The highest concentrations (4.8–7.3 mg/l) occurred in brick tea infusions after 6 h of infusion [19].

### 3.2. Trace metals

Measurements showed that all metals had right-skewed concentration distributions. Al and Mn were the most abundant among the measured metals in black tea with 2.76 and 0.43 mg/l median concentrations, respectively. Descriptive statistics and values of the fitted distribution parameters are presented in Table 2. Median concentrations of the trace metals ranged from 0.11 μg/l for Cd to 89 μg/l for Zn. While concentrations of As, Cd, Co, Cu, Ni, and Zn were below the Turkish non-alcoholic beverage [32] and the American bottled water [31] limits, 95% of the samples had Al concentrations higher than the Turkish non-alcoholic beverage limit of 2 mg/l [32]. Furthermore, 24% of measured Mn concentrations exceeded the American bottled water limit of 0.05 mg/l [31]. Kruskal–Wallis test pointed to a significant difference among loose, pot bag, and instant tea samples for only Al (p = 0.025; p > 0.30 for the remaining metals). Mann–Whitney test showed that the difference was significant between loose and instant bag tea (p = 0.007) with median values of 2.90 and 2.39 mg/l, respectively, and between pot bags (median = 2.77 mg/l) and instant bags (p = 0.055) for Al but not for the other metals. Comparison of metal concentrations among the three most consumed brands pointed that Al and Sr concentrations differed at the presumed significance level in this study. The difference was also significant, but at p = 0.10, for Cu; and p > 0.12 for the remaining metals. Further investigation with Mann–Whitney test between tea brands resulted as while median Sr concentrations were different between Caykur (5.02 μg/l) and Dogus (3.12 μg/l), median Al concentrations were different between Caykur (2.96 μg/l) and Lipton (2.56 μg/l), and median Cu concentrations were different between Dogus (9.5 μg/l) and Lipton (14.1 μg/l).

Metal content of black tea sold in Turkey has not been studied extensively. Colak et al. [23] measured Mn, Fe, Cu, Zn, and Ni concentrations in three green tea samples. Average con-
Table 2
Statistics of trace elemental concentrations in black tea

<table>
<thead>
<tr>
<th>Contaminant</th>
<th>Mean (S.E.)</th>
<th>95% CIAM</th>
<th>Median</th>
<th>S.D.</th>
<th>Min</th>
<th>Max</th>
<th>Skewness</th>
<th>Distribution</th>
<th>Parameter values</th>
</tr>
</thead>
<tbody>
<tr>
<td>Fluoride (mg/l)</td>
<td>0.68 (0.03)</td>
<td>0.63–0.74</td>
<td>0.67</td>
<td>0.20</td>
<td>0.34</td>
<td>1.48</td>
<td>1.13</td>
<td>Logistic</td>
<td>Mean = 0.68; scale = 0.11</td>
</tr>
<tr>
<td>Aluminum (mg/l)</td>
<td>2.91 (0.10)</td>
<td>2.70–3.12</td>
<td>2.76</td>
<td>0.72</td>
<td>1.66</td>
<td>5.35</td>
<td>1.35</td>
<td>Lognormal</td>
<td>Mean = 2.91; S.D. = 0.68</td>
</tr>
<tr>
<td>Arsenic (µg/l)</td>
<td>0.21 (0.03)</td>
<td>0.15–0.26</td>
<td>0.14</td>
<td>0.19</td>
<td>0.002</td>
<td>0.73</td>
<td>1.17</td>
<td>Weibull</td>
<td>Scale = 0.22; shape = 1.10</td>
</tr>
<tr>
<td>Cadmium (µg/l)</td>
<td>0.19 (0.03)</td>
<td>0.13–0.24</td>
<td>0.11</td>
<td>0.19</td>
<td>0.02</td>
<td>0.79</td>
<td>1.97</td>
<td>Lognormal</td>
<td>Mean = 0.19; S.D. = 0.21</td>
</tr>
<tr>
<td>Chromium (µg/l)</td>
<td>5.70 (1.05)</td>
<td>3.60–7.81</td>
<td>3.48</td>
<td>0.74</td>
<td>1.58</td>
<td>43.2</td>
<td>3.83</td>
<td>Lognormal</td>
<td>Mean = 5.11; S.D. = 3.78</td>
</tr>
<tr>
<td>Cobalt (µg/l)</td>
<td>0.35 (0.04)</td>
<td>0.27–0.43</td>
<td>0.29</td>
<td>0.28</td>
<td>0.01</td>
<td>1.58</td>
<td>1.94</td>
<td>Weibull</td>
<td>Scale = 0.38; shape = 1.266</td>
</tr>
<tr>
<td>Copper (µg/l)</td>
<td>17.5 (1.86)</td>
<td>13.7–21.2</td>
<td>12.7</td>
<td>3.57</td>
<td>65.4</td>
<td>1.92</td>
<td>1.13</td>
<td>Lognormal</td>
<td>Mean = 17.1; S.D. = 11.6</td>
</tr>
<tr>
<td>Manganese (µg/l)</td>
<td>572 (56.1)</td>
<td>459–684</td>
<td>429</td>
<td>397</td>
<td>188</td>
<td>2105</td>
<td>1.80</td>
<td>Gamma</td>
<td>Location = 187; scale = 339, shape = 1.13</td>
</tr>
<tr>
<td>Nickel (µg/l)</td>
<td>16.6 (1.00)</td>
<td>14.6–18.6</td>
<td>15.4</td>
<td>7.11</td>
<td>9.60</td>
<td>48.8</td>
<td>3.01</td>
<td>Lognormal</td>
<td>Mean = 16.4; S.D. = 5.4</td>
</tr>
<tr>
<td>Strontium (µg/l)</td>
<td>5.47 (0.80)</td>
<td>3.86–7.07</td>
<td>4.60</td>
<td>5.60</td>
<td>0.03</td>
<td>22.9</td>
<td>2.12</td>
<td>Weibull</td>
<td>Scale = 5.52; shape = 0.99</td>
</tr>
<tr>
<td>Zinc (µg/l)</td>
<td>103 (6.72)</td>
<td>90–117</td>
<td>89</td>
<td>47.5</td>
<td>39.2</td>
<td>253</td>
<td>1.30</td>
<td>Lognormal</td>
<td>Mean = 103; S.D. = 46.7</td>
</tr>
</tbody>
</table>

a Standard error.
b Statistics are based on N = 49 for strontium and N = 50 for the remaining.
c Confidence interval about the mean
d Standard deviation.
e Value of location parameter is 0 for Weibull distribution.

Concentrations were 1610, 342, 32, 28, and 26 mg/kg. Cadmium and vanadium content of black tea was determined by Dundar and Saglam [22]. Mean concentrations of two types of tea were reported as 2.79 and 4.39 mg/kg for cadmium, and 0.65 and 2.30 mg/kg for vanadium.

In a review article, Wong et al. [19] reported that Al in tea infusions ranged between 0.7 and 6.0 mg/l. Al concentrations measured in this study fall in this range. Fernandez et al. [30] analyzed elemental tea infusion concentrations in 22 loose and 2 instant bag tea samples for 11 metals. Range of Al concentrations was 1–10 mg/l. Average concentrations for 5-min infusion of 1.5 g of tea in 100 ml of distilled water were 6.0, 0.1, 3.0, 0.08, and 0.2 mg/l for Al, Cu, Mn, Sr, and Zn, respectively. These mean concentrations are substantially higher than the mean levels measured in this study ranging from approximately 2 times for Al and Zn to 13 times for Sr. Some of the difference arises from the mass of tea infused per water volume, which is 50% higher than this study.

3.3. Individual exposure and risks

3.3.1. Fluoride

Exposures to contaminants were calculated as daily intake (mg/day) and daily intake per body weight (mg/(kg day)) which is chronic daily intake (CDI) in this study. CDI is used to calculate the risk levels due to the exposure by employing a no-adverse-effect-level. The USEPA lists [28] a Reference Dose (RfD) for objectionable dental fluorosis as $6 \times 10^{-2}$ mg/(kg day). Cao et al. [18] reported that an increased risk occurs for bone effect at a total intake of 6 mg F/day. Whyte

Table 3
Descriptive statistics for individual exposure and risk

<table>
<thead>
<tr>
<th>Contaminant</th>
<th>Mean (S.E.)</th>
<th>95% CIAM</th>
<th>Median</th>
<th>S.D.</th>
<th>Min</th>
<th>Max</th>
<th>Skewness</th>
<th>Distribution</th>
</tr>
</thead>
<tbody>
<tr>
<td>Fluoride DI (mg/day)</td>
<td>0.34 (0.04)</td>
<td>0.27–0.41</td>
<td>0.25</td>
<td>0.25</td>
<td>0.04</td>
<td>1.13</td>
<td>0.93</td>
<td></td>
</tr>
<tr>
<td>Fluoride CDF (ng/kg day))</td>
<td>5.0 (0.06)</td>
<td>3.9–6.2</td>
<td>3.3</td>
<td>4.1</td>
<td>0.8</td>
<td>15</td>
<td>1.08</td>
<td></td>
</tr>
<tr>
<td>Fluoride HQ (×10³)</td>
<td>84 (10)</td>
<td>65–100</td>
<td>55</td>
<td>68</td>
<td>13</td>
<td>250</td>
<td>1.08</td>
<td></td>
</tr>
<tr>
<td>Aluminum CDI (µg/kg day)</td>
<td>21.4 (2.53)</td>
<td>16.3–26.5</td>
<td>15.6</td>
<td>17.9</td>
<td>4.13</td>
<td>73.2</td>
<td>1.54</td>
<td></td>
</tr>
<tr>
<td>Aluminum HQ (×10³)</td>
<td>21 (2.5)</td>
<td>16–26</td>
<td>16</td>
<td>18</td>
<td>4.1</td>
<td>73</td>
<td>1.54</td>
<td></td>
</tr>
<tr>
<td>Arsenic HQ (×10³)</td>
<td>5.5 (1.0)</td>
<td>3.6–7.5</td>
<td>2.4</td>
<td>6.8</td>
<td>0.1</td>
<td>26</td>
<td>1.59</td>
<td></td>
</tr>
<tr>
<td>Arsenic R (×10⁶)</td>
<td>2.5 (0.4)</td>
<td>1.6–3.4</td>
<td>1.1</td>
<td>3.1</td>
<td>0.02</td>
<td>12</td>
<td>1.59</td>
<td></td>
</tr>
<tr>
<td>Cadmium HQ (×10³)</td>
<td>2.3 (0.36)</td>
<td>1.5–3.0</td>
<td>1.5</td>
<td>2.5</td>
<td>0.1</td>
<td>12</td>
<td>2.23</td>
<td></td>
</tr>
<tr>
<td>Chromium HQ (×10³)</td>
<td>20 (7.8)</td>
<td>4.5–36</td>
<td>6.8</td>
<td>55</td>
<td>1.0</td>
<td>375</td>
<td>5.76</td>
<td></td>
</tr>
<tr>
<td>Manganese HQ (×10³)</td>
<td>28 (4.0)</td>
<td>20–36</td>
<td>18</td>
<td>28</td>
<td>4.0</td>
<td>131</td>
<td>2.24</td>
<td></td>
</tr>
<tr>
<td>Nickel HQ (×10³)</td>
<td>6.0 (0.6)</td>
<td>4.7–7.2</td>
<td>4.4</td>
<td>4.5</td>
<td>1.0</td>
<td>18</td>
<td>1.05</td>
<td></td>
</tr>
<tr>
<td>Strontium HQ (×10³)</td>
<td>6.7 (1.4)</td>
<td>4.0–9.5</td>
<td>2.9</td>
<td>9.6</td>
<td>0.03</td>
<td>50</td>
<td>2.70</td>
<td></td>
</tr>
<tr>
<td>Zinc HQ (×10³)</td>
<td>2.6 (0.3)</td>
<td>1.9–3.2</td>
<td>1.6</td>
<td>2.4</td>
<td>0.4</td>
<td>11</td>
<td>1.71</td>
<td></td>
</tr>
</tbody>
</table>

a Standard error.
b Confidence interval about the mean.
c Standard deviation.
d Daily intake.
e Chronic daily intake.
f Hazard quotient.
et al. [7] considered 10 mg F/day as the threshold level for “pre-clinical skeletal fluorosis.” Descriptive statistics of individual exposure and risk (HQ) estimations, based on the RfD, are shown in Table 3. None of the participants had exposures that may cause a significant risk (max. HQ = 0.25 < 1) for objectionable dental fluorosis, not to mention the bone effect or skeletal fluorosis from F intake by drinking black tea.

3.3.2. Trace metals

Individual exposure and risk estimates for metals are presented in Table 3. CDI for Al ranged from 4 to 73 µg/(kg day). The World Health Organization (WHO) recommends a tolerable weekly intake of 7 mg/kg body weight [33]. Noncarcinogenic risk levels were estimated using this value in place of RfD since no RfD value was listed by the USEPA [28]. The risk estimates showed that there were no significant risks involved in black tea intake as the maximum HQ value was less than one-tenth of the demarcation value. Researchers have concluded that neither dietary intake [34] nor intake from alum treated drinking water [35] is likely to contribute to Alzheimer’s disease in contrast to the link found in an epidemiological study [8]. In this study, average, standard deviation, and maximum individual daily intake values were calculated as 1.4, 1.1, and 5.1 mg/day, respectively. Taking 25 mg/day [34] as daily adult intake of Al from diet, tea would correspond to 6 and 20% at the mean and maximum levels, respectively. Stauber et al. [35] estimated that contribution of tea to the total Al intake was 53% which was comparable to 41% for food in a total dietary intake of 3.2 mg/day. While the mean daily intake estimated in this study corresponds to 44% of that total intake, the maximum value in this study is 1.6 times the total dietary intake of 3.2 mg/day. Fernandez et al. [30], on the other hand, assumed an average daily dietary intake (ADDI) of 5 mg/day. Nonetheless, Al intake from black tea constituted an important portion of the daily total dietary intake, but the risk levels were not significant.

Tea is, also, an important source of Mn. Fernandez et al. [30] estimated that a consumption of 2.95 cups (50-ml)/day would constitute 10% and 18% of the ADDI (4 mg/day) for loose and instant tea, respectively. The mean (0.25 mg/day) and the maximum (1.2 mg/day) daily intake values estimated in this study are 6.25% and 30% of the ADDI, respectively. The HQ values calculated for Mn did not point to significant noncarcinogenic risk values in this study.

In general, one in million (10⁻⁶) is considered as the acceptable carcinogenic risk level in environmental risk assessment [36]. However, this acceptable level may change according to environmental policies and may be as high as 10⁻⁴ [37,38] as in the case of As. Arsenic concentrations measured in this study were at levels that would result in moderately high carcinogenic risk (R) levels. Average and maximum R values were 2.5 × 10⁻⁶
and $1.2 \times 10^{-5}$, respectively, falling between the two demarcation levels. All noncarcinogenic risk levels for As were well below the demarcation value of 1 as in the cases of Cd, Ni, Sr, and Zn.

### 3.4. Population exposure and risks

#### 3.4.1. Fluoride

Monte Carlo simulation was carried out to estimate distributions of exposure and risk for Izmir population using the fitted distributions of the input variables (see Fig. 1 and Table 2) in the exposure and risk equations (see Section 2.5). The resulting frequency histograms from the simulation for fluoride and the best fitting distributions are presented in Fig. 2. Lognormal distribution was the best fitting distribution to all output variables. None of the risks for dental fluorosis, bone effect, and skeletal fluorosis due to tea intake could be considered high enough to cause a concern for the population. Sensitivity analysis revealed that among the three input variables (fluoride concentration, body weight, and daily tea intake), daily tea intake has the largest effect on the outcomes, exposure and risk. The sensitivity level, measured by rank correlation of an input to the output, was 0.89 for daily tea intake, and approximately 0.30 for the remaining two variables.

Therefore, daily tea intake was analyzed according to participant characteristics (age, sex, education level) and habits (source of tea water, tea type, tea brand, daily drinking water intake) on individual level. Daily tea intake (DI) was found not to differ between males and females. Participants were pooled into three age groups, <15, 15–35, and >35 years, with respective median daily tea intake values of 0.075, 0.30, and 0.75 l/day. Kruskal–Wallis test showed that at least one group had a different median, and Mann–Whitney test showed that all differences were significant. Furthermore, correlation between age group and DI variables was significant ($p < 0.001$) with Spearman’s Rho value of 0.54. Participants were pooled into two education levels (primary + middle school and high school or higher) with median DI values of 0.15 and 0.56 l/day, respectively. The difference was significant ($p = 0.032$) with Mann–Whitney test. Statistical tests showed that DI did not differ with source of tea water ($p = 0.11$), tea type ($p = 0.33$), and tea brand ($p = 0.26$). Mann–Whitney test was applied to test the differences between each of the groups in tea type and brand categories. Results showed that the difference was significant only between the brands of Caykur and Lipton with median values of 0.15 and 0.75 l/day, respectively. No relation was observed between daily tea intake and daily drinking water intake, analyzed by simple linear regression on log-transformed values ($r = 0.03$, $p = 0.83$). In summary, people’s exposure to contaminants in tea infusions was increased with age and education level because their tea consumption was increased with increase in these variables.

#### 3.4.2. Trace metals

Simulation results for the three metals with the highest HQ values, and carcinogenic risk for As are presented as frequency distributions in Fig. 3. The figure also includes the best fitting distribution and its parameter values for the four metals. Lognormal distribution fitted the best to all risk distributions. The three metals are placed in descending HQ range in the figure. Ninety-fifth percentile HQ values for Mn, Al, and Cr were 0.11, 0.08, and 0.05, respectively. Therefore, we may conclude that according to population noncarcinogenic risk distributions, the risk levels were not significant for any of the ten metals evaluated. However, carcinogenic risk for arsenic was of concern since both of the median ($1.2 \times 10^{-6}$) and 95th percentile ($1.0 \times 10^{-5}$) $R$ values are $>10^{-6}$; but not unacceptable, $<10^{-4}$.

Results of sensitivity analyses showed that daily tea intake was the most influencing variable on HQ for all of the three metals presented in Fig. 3. Carcinogenic risk, however, was the most sensitive to the As concentrations (Table 4). Results, in terms of environmental health management, imply that controlling both As concentrations and daily tea intake would be effective in risk mitigation efforts.
Our previous research [39] had shown that carcinogenic risk levels due to oral exposure to drinking water arsenic in Izmir were high with $7.9 \times 10^{-5}$, $3.1 \times 10^{-4}$, and $1.4 \times 10^{-3}$ at median, mean, and 95th percentile levels, respectively. Therefore, as in black tea would add to the carcinogenic risk concerns already high for ingestion of drinking water.

4. Conclusion

Fluoride intake from drinking loose, instant bag, or pot bag teas sold in Izmir, Turkey, was not high enough to cause significant risk for dental fluorosis, not to mention the bone effect or skeletal fluorosis. Exposure to none of the evaluated 10 trace metals via tea intake, including aluminum with regards to Alzheimer’s disease, were not at levels to cause concern for noncancerous risks. However, arsenic was a concern, as the estimated carcinogenic risk levels were greater than one in million.

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References


